

1 **Use of dewatered alum sludge as main substrate in treatment reed bed**  
2 **receiving agricultural wastewater: Long-term trial**

3  
4 Y.Q. Zhao <sup>\*</sup>, X.H. Zhao, A.O. Babatunde

5 Centre for Water Resources Research, School of Architecture, Landscape and Civil Engineering,  
6 University College Dublin, Newstead, Belfield, Dublin 4, Ireland

7  
8 

---

\* Author to whom correspondence should be addressed: Tel: +353-1-7163215; Fax: +353-1-7163297;

9 E-mail: yaqian.zhao@ucd.ie

10  
11 **Abstract**

12 This study aims to explore a novel application of dewatered alum sludge cakes (DASC) as the  
13 main medium in a single model reed bed to treat phosphorus-rich animal farm wastewater  
14 under “tidal flow” operation on a long term basis. It is expected that the cakes act as the  
15 carrier for developing biofilm and also serve as adsorbent to enhance phosphorus (P)  
16 immobilization. Results have demonstrated that average removal efficiencies of  $73.3 \pm 15.9\%$   
17 for COD,  $82.9 \pm 12.3\%$  for BOD<sub>5</sub>,  $86.4 \pm 6.0\%$  for RP (reactive P),  $88.6 \pm 7.2\%$  for SRP (soluble  
18 reactive P) and  $77.6 \pm 17.5\%$  for SS can be achieved during the two year’s operation. More  
19 significantly, the “P-adsorption proportion” by DASC in the reed bed is 42% of the overall P  
20 removal. The remaining removal of P may be contributed by the trapping and filtration  
21 process of DASC. Therefore, the lifetime of the DASC in reed bed is reasonably longer than  
22 that determined from the batch isotherm test.

23 **Keywords:** Dewatered alum sludge, phosphorus, reed bed, tidal flow strategy, wastewater  
24 treatment

## 25 **1. Introduction**

26 It has been recognised and generally accepted that urbanization, industrial revolution, and  
27 deficient waste disposal practices have left a legacy of polluted sediments in water  
28 environment including rivers, estuaries, lakes and seas. Appropriate disposal of such resultant  
29 wastes in line with sustainable development remains a great challenge to engineers and  
30 scientists who are called upon to develop solutions that are technically, economically, and  
31 socially sound. Dewatered alum sludge cakes (DASC) refer to the by-product in water  
32 industry from potable water treatment process where aluminium sulphate is employed for the  
33 raw water flocculation. In most countries worldwide including Ireland, the DASC is regarded  
34 as waste and disposed off in landfill. Although a number of attempts have been made to  
35 beneficially reuse such “waste” as raw materials in civil and environmental engineering  
36 (Babatunde and Zhao, 2007), up till now, landfill remains the dominant option. However, it is  
37 noted that DASC differs significantly from other industrial wastes such as pharmaceutical  
38 industrial waste and petroleum industrial waste that may contain high hydrocarbons and toxic  
39 substances. DASC is derived from the residual of raw water which contains mainly turbidity,  
40 colour and humic substances with no toxic substances in most cases, except probably for the  
41 arsenic in some source waters in special circumstances. Two major features of the DASC are:  
42 (1) high content of aluminium ( $29.7 \pm 13.3\%$  by mass (Babatunde and Zhao, 2007)) and (2) its  
43 locally, largely and easily available nature coupled with the fact that it is free of charge. These  
44 make it possible to reuse such kind of sludge as a valuable raw/resourceful material in  
45 wastewater treatment to enhance adsorption and chemical precipitation processes that remove  
46 various pollutants, especially phosphorus (P).

47 On the other hand, constructed wetlands (or commonly known as reed beds) have been  
48 used successfully worldwide as one of the most popular technologies to treat various types of

49 wastewaters (Scholz, 2006; Healy et al., 2007; Vymazal, 2007, Wood et al., 2007; Babatunde  
50 et al., 2008). In particular, reed bed system is considered to be efficient and at the same time  
51 economically and environmentally attractive and sustainable. In the last several years,  
52 extensive studies at University College Dublin, Ireland have been undertaken to develop  
53 novel approaches for the purpose of enhancing pollutant removal in reed bed treatment  
54 systems. These include "tidal flow" operation strategy and DASC-based reed bed system.  
55 These innovative approaches have demonstrated the improved ability of reed bed systems to  
56 enhance oxygen transfer and a high immobilization capacity for P removal from the  
57 wastewater (Sun et al., 2005; Sun et al., 2006; Babatunde et al., 2007; Zhao et al., 2008).

58 Tidal flow operation strategy is a batch wise, fill-and-draw type operation (Green et al.,  
59 1998; Sun et al., 2005; Sun et al., 2006). When a periodic influent feeding and periodic  
60 discharge is applied, the matrices of the reed bed with wastewater enable the bed matrices to  
61 be fully submerged during the filling process. This provides maximum media-wastewater  
62 contact and avoids the problem of poor wastewater distribution often associated with  
63 conventional continuous-flow reed bed systems. Subsequent draining process then allows air  
64 to be drawn from the atmosphere into the bed matrices without mechanical input, thereby  
65 enhancing aeration and stimulating aerobic biological processes to decompose organic  
66 pollutants and ammoniacal-nitrogen in wastewater.

67 The objective of this study was to examine, in controlled laboratory experiments, the  
68 effectiveness of a model reed bed treatment system with DASC as main filter medium treating  
69 a P-rich wastewater. This study focused on the treatment efficiency in a long term operation  
70 of such kind of treatment reed bed, rather than the feasibility study, which has been reported  
71 previously (Zhao et al., 2008). Moreover, a real animal farm yard wastewater was used in this  
72 study to mimic actual application conditions. The results of this study provide insight into  
73 changes over time in treatment efficiency, BOD<sub>5</sub>, P loading rates and the understanding of the

74 operational lifetime regarding the saturation of the DSAC used. It is expected that the current  
75 study will serve as the basis to establish a DSAC-based reed bed system (say multi-stage  
76 system) for such real animal farm yard wastewater treatment.

77

## 78 **2. Materials and Methods**

### 79 **2.1 Materials**

80 The experimental DASC was collected from the mechanical dewatering unit of a water  
81 treatment plant located in South-west Dublin, Ireland. The plant uses aluminum sulphate for a  
82 reservoir water flocculation at a typical dose of 42-60 mg/l and thus the Al ion is the  
83 dominant component in this DASC. Specific characteristics of the DASC that are relevant to  
84 its use as a reed bed substrate and as a reed growth medium have been examined in detail  
85 (Babatunde et al., 2007). The DASC collected was air-dried and ground and sieved to the  
86 particle size less than 2 mm as the main substrate to be used in a laboratory scale reed bed  
87 system. The wastewater studied was collected periodically from the secondary holding tank of  
88 a local animal farm, which includes about 2000 livestock units of sheep, pigs, cattle and  
89 horses. The animal farm effluent is derived from all the activities on the farm and it undergoes  
90 some form of primary sedimentation before being pumped to the secondary holding tank. The  
91 collected wastewater was allowed for further settlement and the supernatant was stored at  $4 \pm 1$   
92 °C environment.

93

### 94 **2.2 Model reed bed set-up and operation**

95 The single model DASC-based reed bed under investigation was set up using a 145 mm  
96 (internal diameter) Pyrex column filled firstly with gravel to 10 cm as support layer followed  
97 by 35 cm in depth of prepared DASC (2.5 kg) as main substrate. *Phragmites australis* were

98 planted on the top of the bed. The supernatant of the collected wastewater (with or without  
99 dilution with tap water) was used as the influent with concentrations of  $213 \pm 127$  mg/l (COD),  
100  $110 \pm 69$  mg/l ( $BOD_5$ ),  $28 \pm 15$  mg/l ( $PO_4^{3-}$ -P),  $72 \pm 66$  mg/l (SS) and  $6.8 \pm 0.4$  (pH). The influent  
101 was loaded onto the reed bed via a peristaltic pump from a feed tank at a daily flow rate of 8L,  
102 which gives a hydraulic loading rate of  $0.5 \text{ m}^3/\text{m}^2 \cdot \text{d}$ . The treatment system was operated using  
103 the tidal flow strategy for over 730 days. The rhythmical filling and draining generated the  
104 tides and this was realised using peristaltic pumps which were controlled by a preset  
105 electronic timer.

106

### 107 **2.3 Analysis**

108 During the testing period, samples of influent and effluent from the model reed bed were  
109 collected and analysed periodically for  $BOD_5$  (using Hach BODTrack apparatus), COD, SS, P  
110 (using Hach DR/2400 spectrophotometer) and pH (WTW, pH 325, Germany). The P analysis  
111 was based on the reaction of orthophosphate in the samples with molybdate in an acid  
112 medium to produce a mixed phosphate/molybdate complex. In particular, the P analysis was  
113 done in two parts: (1) Samples were directly reacted (without filtration) with the reagent and  
114 analysed to determine the reactive P (RP) and (2) Samples were filtered using a  $0.45 \mu\text{m}$   
115 membrane filter and the filtrate was analysed to determine the soluble reactive P (SRP). In  
116 order to monitor biofilm development onto the surface of DASC used in the reed bed, surface  
117 examination/observation of the fresh alum sludge particles (referred to as clean sludge) and  
118 the sludge particles that were used in the reed bed system (referred to as used sludge) were  
119 carried out under a scanning electron microscope (SEM) using a JEOL JSM 6400 Scanning  
120 Electron Microscope. In addition, measurement of molecular size distribution (MSD) of  
121 dissolved organic substances using a high-pressure size exclusion chromatography (HPSEC)  
122 for the samples of influent and effluent at 705<sup>th</sup> day was carried out for the purpose of

123 providing insight into the pollutants removal inside the reed bed. HPSEC consists of a Waters  
124 1515 isocratic pump, a Waters 2487 UV dual  $\lambda$  detector operated at 254 nm and a PL  
125 Aquagel-OH 40 (300×7.5 mm) column. Molecular weight standards were composed of  
126 sodium polystyrenesulfonates (35, 18, 8, 5.4, and 1.8K) and acetone.

127

### 128 **3. Results**

129 Fig. 1 illustrates the progressive treatment performance of the reed bed. From the results, the  
130 initial removal of the carbonaceous substrates averaged 70% for BOD<sub>5</sub> and 56% for COD  
131 although this would mainly be attributable to filtration due to the lack of biological activities  
132 in the newly setup reed bed matrix. However, with the gradual establishment of a dynamic  
133 biological system through intense activities of the reeds and microorganisms in the alum  
134 sludge matrix, improved removal of over 80% for both BOD<sub>5</sub> and COD was observed after  
135 100 day's operation. Additionally, enhanced oxygen supply due to the tidal flow operation  
136 strategy led to over 90% BOD<sub>5</sub> removal by the 260<sup>th</sup> day with a short period of exception as  
137 shown in Fig. 1. The calculated value of the theoretical oxygen supply rate by the tidal flow  
138 operation strategy used in the system is 137.2g/m<sup>2</sup>.d (Babatunde, 2007). The average removal  
139 of 73.3±15.9% for COD, 82.9±12.3% for BOD<sub>5</sub> was maintained during the entire  
140 experimental period. It can be seen from Fig. 1 that the COD removal exhibited some  
141 considerable fluctuations, and this may be caused by integrated removal of enhanced  
142 adsorption, filtration and biological degradation of pollutants in terms of COD. Relatively,  
143 BOD<sub>5</sub> removal was solely from the biological degradation. The relative low removal of BOD<sub>5</sub>  
144 and COD of the reed bed in the later stage of the operation period can be adduced to  
145 operational reasons as the laboratory was relocated during that period and this affected the  
146 routine operation.

147

[Fig. 1 here]

148

149 Fig. 2 provides evidence of a good correlation between BOD<sub>5</sub> loading and its removal (in  
150 g/m<sup>2</sup>.d). Increased BOD<sub>5</sub> loading resulted in higher BOD<sub>5</sub> removal in terms of g/m<sup>2</sup>.d of the  
151 reed bed, thus indicating the intensive activities of the microorganisms inside the reed bed.  
152 Data of pH monitoring were not presented since there were no specific characteristics related  
153 to the pH except for the trend of effluent pH being lower than that of influent due to the  
154 hydrolysis of DASC in the bed (Yang et al., 2006).

155 [Fig. 2 here]

156

157 Removals of both RP and SRP were significant in the reed bed throughout the experiments.  
158 The average removal efficiencies of 86.4±6.0 % for RP and 88.6±7.2 % for SRP, respectively,  
159 were achieved under the P-loading of 13.5±7.2 g-PO<sub>4</sub><sup>3-</sup>/m<sup>2</sup> d for RP and 8.2±3.2 g-PO<sub>4</sub><sup>3-</sup>/m<sup>2</sup> d  
160 for SRP, respectively. Mass balance estimation shows that about 35.8 mg-PO<sub>4</sub><sup>3-</sup>/g-sludge was  
161 adsorbed by DASC in the reed bed during the operation. Interestingly, the DASC in the bed  
162 has not been saturated and a P removal efficiency of 70 % was still obtained, although an  
163 obvious decline of removal efficiency is observed from Fig. 3. In-depth study has revealed  
164 that the main P-adsorption mechanism is ligand exchange reactions (Yang et al., 2006). This  
165 highlights the obvious advantage of the use of the DASC in enhancing P removal in the  
166 system.

167 [Fig. 3 here]

168

169 Behaviour of SS removal in the reed bed is shown in Fig. 4. Although an average SS removal  
170 of 77.6±17.5% was obtained, and the DASC has been demonstrated as a potential medium in  
171 reed bed for trapping and filtering SS, it has to be stressed that there was ample fluctuation in  
172 the SS removal, and therefore, necessary care should be taken to further study this issue.

173 During the reed bed operation period, serious problems, such as ponding, did not occur, but  
174 concerns of bed clogging in relation to proper arrangement of DASC in the reed and the pre-  
175 treatment of the raw wastewater still warrant further investigation.

176 [Fig. 4 here]

177

## 178 **4. Discussion**

### 179 **4.1 BOD<sub>5</sub> and COD removal**

180 The goal of this study lies in the long term examination of DASC which is expected to act as  
181 a carrier for developing biofilm and also serve as adsorbent to enhance P immobilization from  
182 a real wastewater treatment. As a biofilm carrier, the DASC has successfully demonstrated its  
183 role. The significant removals of both BOD<sub>5</sub> and COD (see Fig. 1) have provided the vital  
184 evidence of biofilm development and intensive biological activities inside the reed bed. In  
185 addition, both the measurements of MSD and SEM have demonstrated the development of  
186 biofilm in the reed bed and the role of DASC as the biofilm carrier although the MSD data  
187 and the SEM image are not shown. MSD measurement provides the evidence which differs  
188 from COD and BOD<sub>5</sub> measurement to illustrate the molecular size distribution of dissolved  
189 organic pollutants before and after the reed bed treatment. The result support that the reed bed  
190 degraded/broke the organics from large molecular sizes into smaller ones via mainly  
191 biological activities inside the bed medium. SEM observation of clear and used sludge  
192 samples showed an obvious difference in morphology. The used sludge was fully covered by  
193 thick slimes, which showed fluffy and transparent characteristics. While clear sludge showed  
194 a crystal lattice structure surface. The striking change of the surface characteristics may  
195 provide the evidence of biofilm development onto the sludge surface, indicating the growth  
196 and considerable activities of the microorganisms although the SEM here is a qualitative tool  
197 to describe the biofilm formation.

198 The influent and effluent of COD and BOD<sub>5</sub> during the testing period are illustrated in Fig.  
199 5. Continuous reduction of COD and specially BOD<sub>5</sub> provides a definite indication of  
200 biological degradation of the organics from the wastewater. It should be noted that except for  
201 the biological process inside the reed bed, a multi-function of the bed which involves  
202 adsorption, filtration, precipitation processes is existed in parallelism. The removal of COD  
203 and BOD<sub>5</sub> is the results of the integrated functions inside the bed. Therefore, it is suggested  
204 that the monitoring of such DASC-based reed bed treatment system should use both COD and  
205 BOD<sub>5</sub> as benchmarks to reveal the degree of treatment efficiency, rather than using one  
206 (either COD or BOD<sub>5</sub>) as index which has been frequently employed in wastewater treatment  
207 practice. In addition, considering the much strong animal farm wastewater in practice, it  
208 should be pointed out that a single reed bed is impossible to satisfy with the treatment  
209 standard. A multi-stage reed bed system is thus suggested and further tests should be  
210 conducted. However, it is out of the scope of the current study.

211

212 [Fig. 5 here]

213

#### 214 **4.2 Phosphorus removal and the lifetime of DASC in reed bed**

215 During the monitoring period, the DASC exhibited good ability as adsorbent to achieve high  
216 P immobilization (see Fig. 3). Both RP and SRP were monitored in this study and there was a  
217 good correlation between them as illustrated in Fig. 6, suggesting that either RP or SRP can  
218 be used as sole index for P removal in practice for such kind of reed bed treatment system.  
219 Fig. 7 shows the RP and SRP removal (in g/m<sup>2</sup>.d) with the RP loading rate applied to the reed  
220 bed studied. It is noted that although the RP loading rates varied 6.2-32.2 g PO<sub>4</sub><sup>3-</sup>-P/m<sup>2</sup>.d, the  
221 reed bed system could remove P efficiently even as the loading rates increased, as indicated  
222 by a linear relationship between P loading and removal. This suggests a good potential and

223 high P adsorption ability of the DASC. From the linear regression analysis (see Fig. 7), it is  
224 interesting to note that the slopes of the two linear regressions are obviously different. If the  
225 removal of SRP refers mainly to the adsorption and biological assimilation of the  
226 microorganisms growth, then the gap between the two lines could be considered as an  
227 indication of the sludge's 'other' functions of physical trapping, filtration and sedimentation  
228 in the bed. Obviously, adsorption plays 42% (0.38/0.90) role in P removal based on the data  
229 of the current study. Thus it is reasonable to state that the maximum adsorption capacity  
230 obtained experimentally from batch isotherms tests cannot be used as a guide to the actual  
231 lifetime of such reed bed in real operation, although attempts have been made to do so in  
232 literature with little or no success (Arias and Brix, 2005; Dong et al., 2005; Park and  
233 Polprasert, 2008; Zhao et al., 2008). It is understood that the batch isotherm tests is solely  
234 designed for the determination of adsorption, eliminating filtration, sedimentation and  
235 biological functions of the material tested. Therefore, the lifetime of the DASC regarding its P  
236 adsorption saturation in reed bed could be much longer than that determined based on batch  
237 isotherm test.

238 [Fig. 6 here]

239 [Fig. 7 here]

240

### 241 **4.3 The proactive nature of the DASC-based reed bed treatment system**

242 The success of this innovative reed bed has demonstrated the beneficial and sustainable reuse  
243 of DASC (a hitherto designated waste product) as a raw material in wastewater treatment  
244 engineering. Regarding its large scale application in practice, some important issues such as  
245 possible clogging of the bed and release of some substances from the DASC still need to be  
246 studied and properly addressed. Unlike other waste by-products derived from various  
247 industries, DASC remains an inescapable by-product from the current drinking water

248 treatment process/technology. The continuous supply of such by-product is certain for now in  
249 water treatment industry and it is free of charge. For example, only in Ireland, a double-fold  
250 increase in alum sludge generation has been forecast by the end of the next decade, from a  
251 current estimate of 15,000–18,000 t/Pa of the DASC (Yang et al., 2006). Application of  
252 DASC in reed bed system will further reduce the capital invest of the reed bed treatment  
253 system and make such treatment technology more attractive. From a previous study by Zhao  
254 et al. (2007) using the same DASC as used in the current study, the P adsorption capacity of  
255 the DASC determined using batch isotherm tests would imply that the lifetime of the DASC  
256 used in the reed bed for domestic wastewater treatment could be 4-17 years based on the  
257 estimate described by Xu et al., (2006). By considering the “adsorption proportion” of 42% in  
258 overall P removal derived from the current study, the lifetime of the DASC can be 9-40 years.  
259 Even in the case of high P effluent of animal farm wastewater treatment tested in this study,  
260 the lifetime of DASC can be estimated as 2.5-3.7 years. In most cases of the practical use of  
261 reed bed system, a multi-stage treatment system is usually applied. Therefore, the lifetime is  
262 reasonably expected to be longer. However, as pointed out by Xu et al., (2006), such an  
263 estimate of the lifetime of a reed bed should only be used as a guide because the P sorption  
264 capacity is influenced by many factors. The most examined factors are particle size, pH,  
265 initial P concentration and temperature etc (Ippolito et al., 2003; Dayton and Basta, 2005;  
266 Yang et al., 2006; Zhao et al., 2007; Korkusuz et al., 2007). For this reason, the P adsorption  
267 capacity could vary in magnitude (Korkusuz et al., 2007). Nevertheless, DASC-based reed  
268 bed has clearly demonstrated its advantage and is expected to be a promising idea in  
269 constructed wetland technology for wastewater treatment.

270 It should be pointed out that two major issues related to the potential application of such  
271 DASC-based reed bed have not been fully studied. One is the clogging tendency of such  
272 sludge reed bed since the SS trapped in the bed increases the average water retention time and

273 thus reduces the effective area available for water flow. Reduced hydraulic conductivity and  
274 the infiltration rate lead to ponding, which represents a serious threat to the normal operation  
275 for the treatment function. Although obvious bed clogging did not occur in this study, the  
276 scatter of the SS removal efficiency (see Fig. 4) indicates that SS removal should be specially  
277 studied to avoid bed clogging. One suggestion could be made that the use of the DASC-based  
278 reed bed as a stand-alone treatment system would not be recommended and pre-treatment to  
279 remove SS would be necessary and vital to result in the successful operation. The other issue  
280 is the possible release of some substances from DASC including polymer residual to the  
281 effluent from DASC since polymer is normally added as sludge conditioner to enhance its  
282 dewaterability. Polymer residual in the DASC and its potential release when the sludge is  
283 reused should be studied since the toxicity of degraded polymer in the environment remains  
284 an unknown health risk from the long term point of view (Majam and Thompson, 2006).

285

## 286 **5. Conclusions**

287 Two years experiment was conducted in a model DASC-based reed bed aimed to exploring  
288 the feasibility and effectiveness regarding P-rich wastewater treatment. The novel use of  
289 DASC as the main reed bed substrate is a promising idea in constructed wetland technology.  
290 The average removal efficiencies of  $73.3 \pm 15.9\%$  for COD,  $82.9 \pm 12.3\%$  for BOD<sub>5</sub>,  $86.4 \pm 6.0\%$   
291 for RP,  $88.6 \pm 7.2\%$  for SRP and  $77.6 \pm 17.5\%$  for SS were achieved during the entire operation  
292 period. Results revealed that the “P-adsorption proportion” by DASC in the reed bed is 42%  
293 of overall P removal. Therefore, the lifetime of the DASC is reasonably longer than that  
294 determined from batch isotherm test. Regarding the large scale application of such reed bed  
295 system, further studies on bed clogging tendency and possible release of some substances  
296 inside the DASC are recommended.

297

298 **Acknowledgements**

299 Authors are indebted to the Irish Environmental Protection Agency for providing financial  
300 assistance for this study through the Environmental Technologies Scheme (project No. 2005-  
301 ET-MS-38-M3). The second author would like to thank University College Dublin for the  
302 financial support granted through *Ad Astra* scholarship. The Lyons Estate Farm is also  
303 sincerely thanked for its cooperation.

304

305

306

307 **References**

308 Arias, C.A., Brix, H., 2005. Phosphorus removal in constructed wetlands: can suitable  
309 alternative media be identified? *Water Sci. Technol.* 51(9), 267–273.

310 Babatunde, A.O., 2007. The development of an alum sludge based tidal flow constructed  
311 wetland for optimizing phosphorus and organic matter removal from wastewaters. PhD  
312 dissertation, University College Dublin, Ireland.

313 Babatunde, A.O., Zhao, Y.Q., 2007. Constructive approaches towards water treatment works  
314 sludge management: An international review of beneficial re-uses, *Critical Reviews in*  
315 *Environmental Science and Technology.* 37(2), 129-164.

316 Babatunde, A.O., Zhao, Y.Q., O'Neill, M., O'Sullivan, B., 2008. Constructed wetlands for  
317 environmental pollution control: A review of developments, research and practice in  
318 Ireland. *Environment International.* 34, 116-126.

319 Babatunde, A.O., Zhao, Y.Q., Yang, Y., Kearney, P., 2007. From 'fills' to filter: Insights into  
320 the reuse of dewatered alum sludge as a filter media in a constructed wetland. *Journal of*  
321 *Residuals Science & Technology.* 4, 147-152.

322 Dayton, E.A., Basta, N.T., 2005. A method for determining the phosphorus sorption capacity  
323 and amorphous aluminum of aluminum-based drinking water treatment residuals. *J.*  
324 *Environ. Qual.* 34, 1112–1118.

325 Dong, C.S., Ju, S.C., Hong, J.L., Jong, S.H., 2005. Phosphorus retention capacity of filter  
326 media for estimating the longevity of constructed wetland. *Water Res.* 39, 2445–2457.

327 Green, M., Fiedler, E., Safrai, I., 1998. Enhancing nitrification in vertical flow constructed  
328 wetland utilizing a passive air pump. *Water Res.* 32, 3513-3520.

329 Healy, M.G., Rodgers, M., Malqueen, J., 2007. Treatment of dairy wastewater using  
330 constructed wetlands and intermittent sand filters. *Bioresour. Technol.* 98, 2268–81.

331 Ippolito, J.A., Barbarick, K.A., Heil, D.M., Chandler, J.P., Redente, E.F., 2003. Phosphorus  
332 retention mechanisms of a water treatment residual. *J. Environ. Qual.* 32, 1857–1864.

333 Korkusuz, E.A., Beklioglu, M., Demirer, G.N., 2007. Use of blast furnace granulated slag as a  
334 substrate in vertical flow reed beds: Field application. *Bioresour. Technol.* 98, 2089–2101.

335 Majam, S., Thompson, P.A., 2006. Polyelectrolyte determination in drinking water. *Water*  
336 *SA*, 32, 705-707

337 Park, W.H., Polprasert, C., 2008. Phosphorus adsorption characteristics of oyster shells and  
338 alum sludge and their application for nutrient control in constructed wetland system.  
339 *Journal of Environmental Science and Health(A)*, 43, 511-517.

340 Scholz, M., 2006. *Wetland systems to control urban runoff*, Elsevier, Amsterdam, The  
341 Netherlands.

342 Sun, G., Zhao, Y.Q., Allen, S.J., 2005. Enhanced removal of organic matter and ammoniacal-  
343 nitrogen in a column experiment of tidal flow constructed wetland system. *Journal of*  
344 *Biotechnology.* 115, 189-197.

345 Sun, G., Zhao, Y.Q., Allen, S.J., Cooper, D., 2006. Generating “Tide” in pilot-scale  
346 constructed wetlands to enhance agricultural wastewater treatment, *Engineering in Life*  
347 *Sciences*. 6, 560-565.

348 Vymazal, J., 2007. Removal of nutrients in various types of constructed wetlands, *Science of*  
349 *The Total Environment*. 380, 48-65.

350 Wood, J., Fernandez, G., Barker, A., Gregory, J., Cumby, T., 2007. Efficiency of reed beds in  
351 treating dairy wastewater. *Biosystems Engineering*, 98, 455 – 469.

352 Xu, D., Xu, J., Wu, J., Muhammad, A., 2006. Studies on the phosphorus sorption capacity of  
353 substrates used in constructed wetland systems. *Chemosphere*, 63, 344–352.

354 Yang, Y., Zhao, Y.Q., Babatunde, A.O., Wang, L., Ren, Y.X., Han, Y., 2006. Characteristics  
355 and mechanisms of phosphate adsorption on dewatered alum sludge. *Sep. Pur. Technol.*  
356 51, 193-200.

357 Zhao, Y.Q., Babatunde, A.O., Razali, M., Harthy, F., 2008. Use of dewatered alum sludge as  
358 a substrate in reed bed treatment systems for wastewater treatment. *Journal of*  
359 *Environmental Science and Health(A)*, 43, 105-110.

360 Zhao, Y.Q, Razali, M., Babatunde, A.O., Yang, Y., Bruen, M., 2007. Reuse of Aluminium-  
361 based water treatment sludge to immobilize a wide range of phosphorus contamination:  
362 Equilibrium study with different isotherm models. *Separation Science and Technology*, 42,  
363 2705-2721.

364

365

366

367

368

369

370

371 **Figure caption:**

372

373 Figure 1. Removal efficiencies of COD and BOD<sub>5</sub> throughout the experimental period

374 Figure 2. Correlation between BOD<sub>5</sub> loading and its removal

375 Figure 3. Removal efficiencies of RP and SRP throughout the experimental period

376 Figure 4. Removal efficiencies of SS throughout the experimental period

377 Figure 5. Variations of COD and BOD<sub>5</sub> throughout the experimental period

378 Figure 6. Correlation between RP and SRP removals

379 Figure 7. Correlation between RP loading and the removals (g/m<sup>2</sup>.d) of RP and SRP

380

381

382

383

384

385

386

387

388

389

390

391

392

393

394

395

396

397

398

399

400

401

402

403

404

405

406

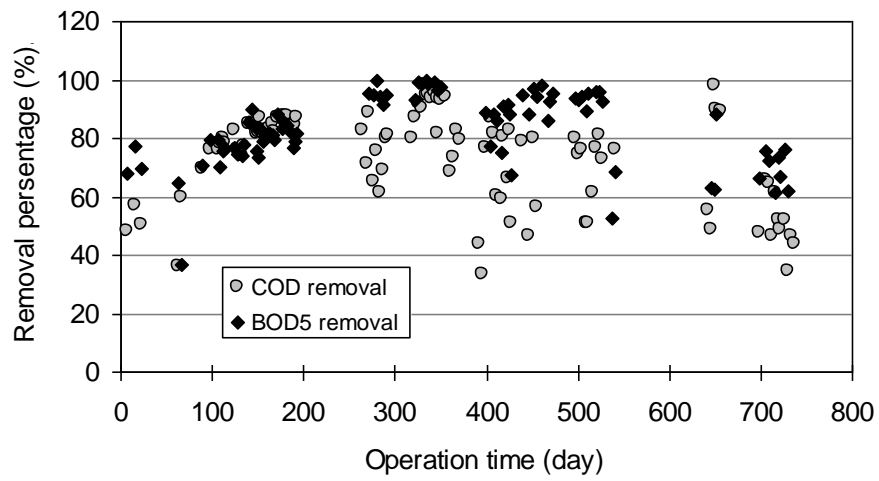
407

408

409

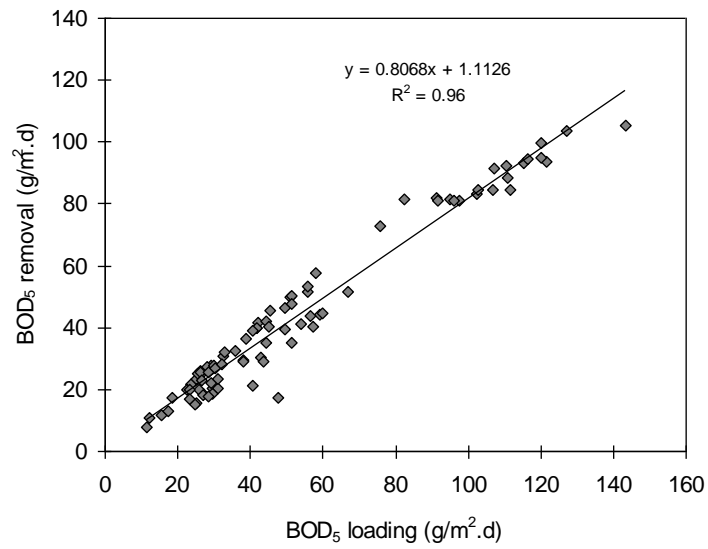
410

411



412  
413  
414  
415  
416  
417  
418  
419

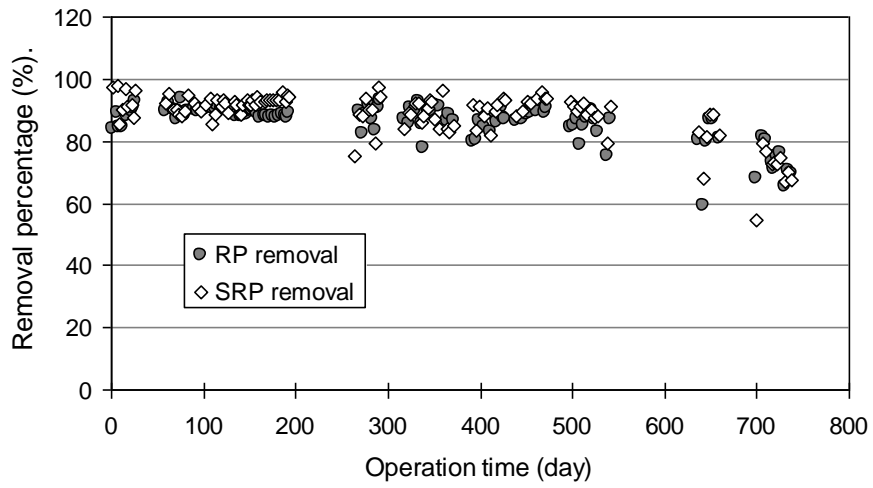
Fig. 1



420  
421  
422  
423  
424  
425  
426  
427  
428  
429  
430

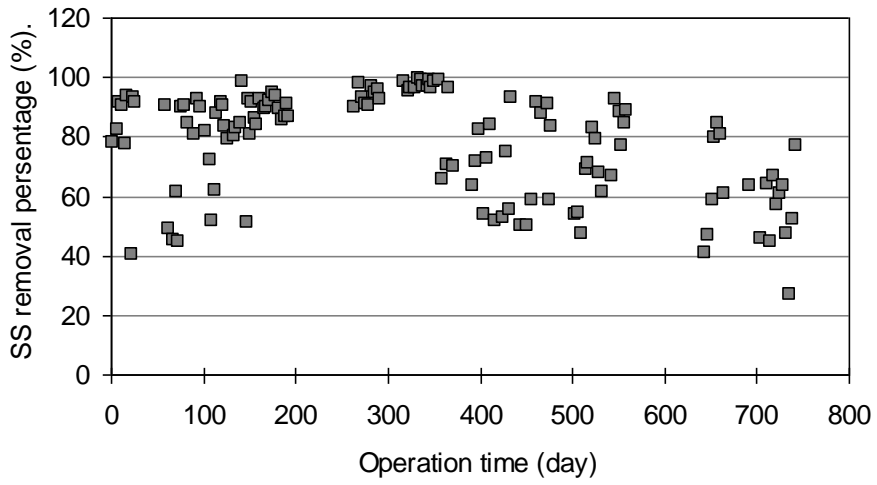
Fig. 2

431



432  
433  
434  
435  
436  
437  
438

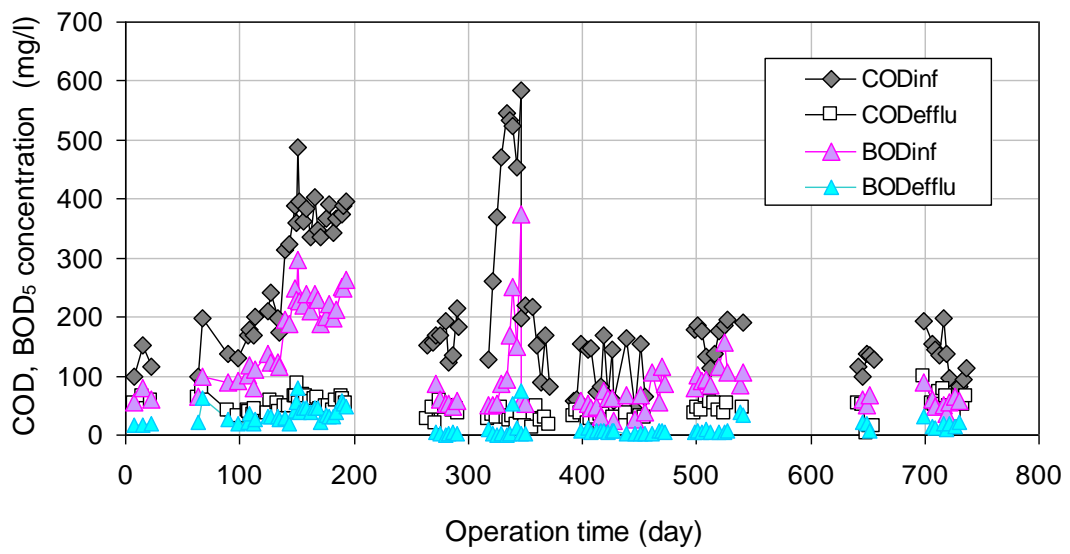
Fig. 3



439  
440  
441  
442  
443  
444  
445  
446  
447  
448  
449  
450  
451

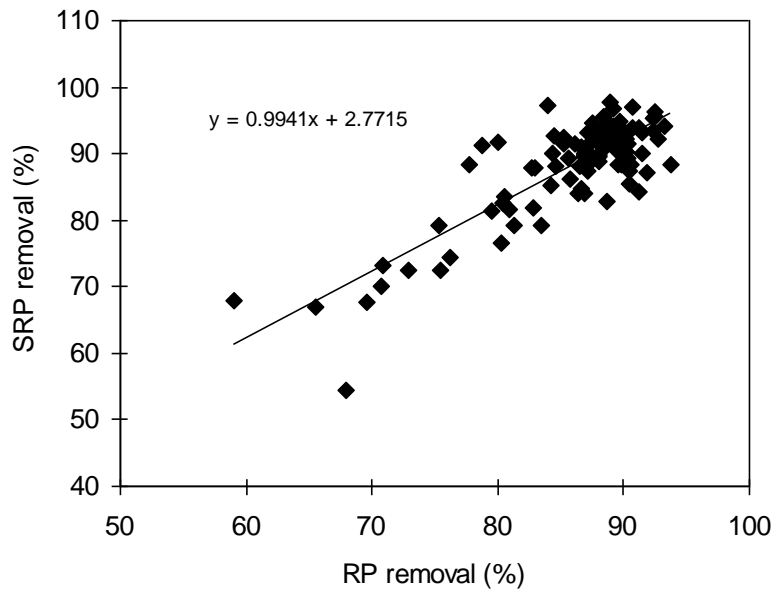
Fig. 4

452



453  
454  
455  
456  
457  
458  
459  
460  
461  
462  
463

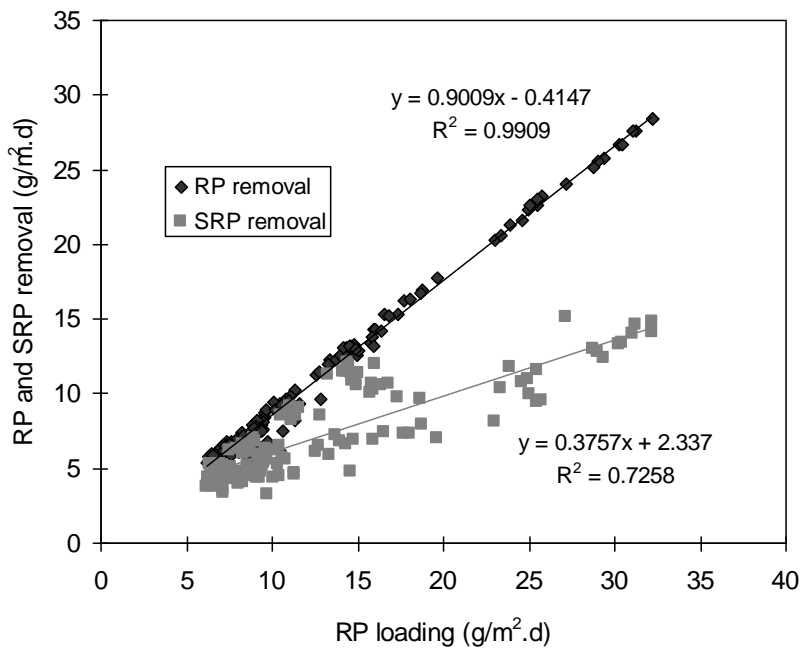
Fig. 5



464  
465  
466  
467  
468

Fig. 6

469  
470



471  
472  
473  
474  
475

Fig. 7