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1	Simultaneous removal of malachite green and hexavalent chromium by
2	Cunninghamella elegans biofilm in a semi-continuous system
3	
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14 Abstract

- 15 The present study was conducted to evaluate the potential of the fungus *Cunninghamella*
- 16 *elegans* for simultaneous decolourisation of a triphenylmethane dye malachite green (MG)
- 17 and hexavalent chromium [Cr(VI)] in the same media. This fungus can degrade MG through
- 18 its reduction into leucomalachite green and then demethylation followed by oxidative
- 19 cleavage. Along with MG degradation, *C. elegans* biofilm could effectively and repeatedly
- remove Cr(VI) from the liquid cultures even in the presence of high concentrations (40 g L⁻¹)
- of NaCl and various other metal ions. *C. elegans* biofilm was also found to adsorb different
- dyes (reactive black-5, acid orange 7, direct red 81 and brilliant blue G) concurrently with
- 23 Cr(VI). Based on its potential for simultaneous removal of dyes and Cr(VI) as well as
- reusability, *C. elegans* biofilm is envisaged as an efficient bioresource to devise strategies for
- 25 treatment of wastewaters loaded with multiple pollutants.
- 26
- 27 Keywords: Immobilization; fungus; dye decolorisation; textile wastewater; adsorption
- 28

29 **1. Introduction**

Dyes are a common constituent of wastewaters originating from various industrial processes. 30 Malachite green (MG) is a triphenylmethane cationic dye which is used in textile, leather, 31 medical, food and paper industries in addition to its use as a biocide to control protozoan and 32 fungal infections in fish farming (Culp and Beland, 1996; Srivastava et al., 2004). However, 33 34 discharge of MG-loaded wastewaters into the environment reduces light penetration in the water bodies and affects the living organisms present owing to the carcinogenic, mutagenic 35 and teratogenic properties of MG and its metabolites (Culp and Beland, 1996; Srivastava et 36 37 al., 2004; Donya et al., 2012). For example, MG is toxic to mammalian cells and has been shown to cause cancer in different organs including liver and thyroid of experimental animals 38 (Rao, 1995; Srivastava et al., 2004; Donya et al., 2012). Leucomalachite green, which is a 39 major metabolite arising from the reduction of malachite green, is also of particular concern 40 owing to its toxicity, mutagenicity and its relatively higher lipophilicity, which result in it 41 42 being retained in fish muscle and fat (Bilandzic et al., 2012). Despite the fact that MG has been banned in some countries it is still being used in others owing to its low cost, ready 43 44 availability and high efficacy. In addition to dyes, wastewaters originating from different industries, including textile and leather, have also been found to contain considerable 45 46 amounts of different salts and metal ions (Tuzen et al., 2008; Ngah and Hanafiah, 2008). The latter are present either from the use of metal complex dyes or metal-containing salts as 47 mordant for better fixation of dyes. Among the metal ions, hexavalent chromium [Cr(VI)] is a 48 common pollutant which co-exists with dyes in the wastewaters originating from textile and 49 50 leather industries (Desai et al., 2009). It is not only the second most common inorganic contaminant of ground water and hazardous waste sites but also listed by the United States 51 52 Environmental Protection Agency among the 17 chemicals for posing the greatest threat to human health (Horton et al., 2006; Cheung and Gu, 2007; Quintelas et al., 2008). In addition 53 54 to disruption of biochemical and physiological functions in bio-systems owing to its strong oxidizing nature, high solubility in water and rapid permeability, it has also been reported to 55 56 harbor mutagenic, carcinogenic and teratogenic properties (McLean and Beveridge, 2001; Ilias et al., 2011). Hence, the co-existence of Cr(VI) and synthetic dyes, including malachite 57 58 green, in wastewaters is a matter of serious concern and there is a need to find effective, innovative and economic treatment technologies to eliminate them or minimize their quantity 59 in the environment. 60

Exploitation of microorganisms for bioremediation of contaminated environments has
 attracted attention as a cost-effective and environmentally friendly approach. Several

researchers have isolated and characterized various bacterial and fungal strains for removal 63 and detoxification of chromium in soil and water resources (Prigione et al., 2008; Dhal et al., 64 2010; Ilias et al., 2011; Essahale et al., 2012; Maqbool et al., 2015). Similarly, a number of 65 bacterial strains belonging to different genera have been isolated and characterized for 66 decolourisation of MG (Li et al., 2009; Kalyani et al., 2012). The potential for 67 68 decolourisation and degradation of this dye has also been reported in various fungi including 69 Phanerochaete chrysosporium, Cyathus bulleri, Cyathus stercoreus, Cyathus striatus, and Penicillium ochrochloron (Vasdev et al., 1995; Jadhav and Govindwar, 2006; Shedbalkar and 70 71 Jadhav, 2011; Jasinska et al., 2012). The non-lignolytic fungus Cunninghamella elegans is well known for its ability to transform a broad range of xenobiotics (Murphy, 2015) and the 72 inactivated biomass of the fungus is an effective biosorbent (Tigini et al., 2010). Cha et al. 73 (2001) observed the formation of leucomalachite green, N-demethylated and N-oxidized 74 metabolites upon incubation of C. elegans with the MG. Microsomal fractions also catalysed 75 76 the production of leucomalachite green and N-demethylated metabolites, and the 77 biotransformation was inhibited by 1-aminobenzotriazole, metyrapone and SKF 525-A, thus 78 it was reasoned that the reduction and *N*-demethylation reactions were catalysed by cytochrome P450. Kim et al. (2010) purified a cytochrome c, CeCyt, from the mitochrondria 79 80 of C. elegans, that catalysed the decolourisation of malachite green and suggested that the protein functions to reduce malachite green under conditions of oxidative stress. 81 82 Whilst there are some reports on the simultaneous removal of different dyes and Cr(VI) from synthetic textile wastewaters by using some multifunctional bacterial strains 83 84 (Desai et al., 2009; Mahmood et al., 2013; Anwar et al., 2014; Maqbool et al., 2016), to the best of our knowledge, simultaneous microbial removal of MG and Cr(VI) has not yet been 85 the focus of any study. Moreover, there is no report regarding the application of fungal strains 86 for such simultaneous removal of dyes and metal ions. In this context, the present study has 87 88 been conducted for simultaneous removal of MG and Cr(VI) by using C. elegans. Biofilms of this fungus have already been reported to demonstrate improved biotransformation of 89 drugs and xenobiotics compared with suspended cells (Amadio et al., 2013; Mitra et al., 90 2013; Quinn et al., 2015). The aim of this study is to extend the possible application of the 91 92 fungal biofilm to the bioremediation of dye/metal contaminated wastewater. 93 2. Materials and Methods 94

95 **2.1 Dyes**

- Malachite green (technical grade) was acquired from BDH (Poole, UK), reactive black-5 and
 direct red-81 were purchased from Santa Cruz Biotechnology (Dallas, TX, USA), and acid
 orange-7 (≥ 85 %) and brilliant blue G were obtained from Sigma Aldrich (Arklow, Ireland).
- 99

100 2.2. Cultivation of *C. elegans* biofilm and planktonic cells

Cunninghamella elegans DSM 1908 was grown on sabouraud glucose agar for 120 h at 28 101 102 °C. Inoculum was prepared by homogenizing one plate of agar and mycelia in 100 mL of 0.8% autoclaved saline. The planktonic cell cultures were grown in 250 mL Erlenmeyer 103 104 flasks containing 45 mL of sterilised sabouraud dextrose broth and 5 mL of C. elegans homogenate. For cultivating biofilms the method described by Amadio et al. (2013) was 105 followed. For biofilm cultivation, stainless steel compression springs (1.2 mm, T316 wire, 106 Shannon Coiled Springs, Ireland) were placed at the bottom of 250 mL Erlenmeyer flasks 107 containing sterilized sabouraud dextrose broth (49 mL). The springs were kept completely in 108 109 contact with the inner walls of the flasks for optimum biofilm growth. Each flask was inoculated with C. elegans homogenate (1 mL) and incubated for 72 h with rotary agitation 110 (150 rpm) at 28 °C. 111

112

113 2.3. Decolourisation of MG by *C. elegans* biofilm and planktonic cells

After 72 h of biofilm growth, the medium in the flasks was replaced with 50 mL sterile MG 114 aqueous solution (80 μ M or 29 mg L⁻¹, unless stated) and incubated with shaking (150 rpm) 115 at 28 °C, alongside an un-inoculated control. When the MG was degraded and both the 116 supernatant and biofilm had been decolourized, fresh MG was added to the flasks. For 117 estimation of decolourisation by planktonic cultures, the cells were harvested by centrifuging 118 (3500 rpm for 15 min) and the biomass was re-suspended in 50 mL of the aqueous MG 119 solution and incubated as before; for the biofilm cultures, the supernatant was decanted and 120 replaced. The supernatants (1.5 mL) and biomass (200 mg) of biofilm and planktonic cells 121 were collected aseptically at regular intervals. Malachite green decolourisation in the 122 supernatant was determined spectrophotometrically as previously described (Jasinska et al., 123 2012) by measuring the change in absorbance at 617 nm (λ_{max}). The biomass was immersed 124 125 in 1 mL of methanol and shaken vigorously for 30 s. The methanol extract was then used to determine malachite green decolourisation (Nanodrop 1000). To recycle biofilms, the 126 supernatants were decanted directly and fresh aqueous dye solution was added. Planktonic 127 cells were harvested by centrifuging and the biomass re-suspended in 50 mL of fresh dye. 128

- 129 The biofilms and the planktonic cells were both rejuvenated by replacing the supernatants
- 130 with 50 mL of fresh sabouraud dextrose broth and incubated for up to 16 h.
- 131

132 2.3.1. Effect of pH on MG decolourisation by *C. elegans* biofilm

- To determine the effect of pH on MG decolourisation, biofilm was cultivated as describedand incubated for 48 h with dye dissolved in water (50 mL). The supernatant was decanted
- and replaced by 50 mL of 20 mM phosphate buffer (pH 3, 5-7) or 2-(*N*-morpholino)
- ethanesulfonic acid (pH 4) containing malachite green; the pH experiments were conducted
- 137 with the same biofilm, starting with pH 7 and ending at pH 3.
- 138

139 **2.3.2. Metabolite identification**

The degradation products of malachite green were extracted from the biomass by incubating
it with 50 mL of ethyl acetate for 3 hours. The organic layer was evaporated to dryness, the

- residue redissolved in 1 mL ethyl acetate and analysed by gas chromatography-mass
- spectrometry (GC-MS) using a method similar to that described by Du et al. (2011). Samples
- 144 (1 μ L) of the extract were injected in the splitless mode onto a HP5MS column (30 m \times 0.25
- 145 mm \times 0.25 µm). The oven temperature held at 120 °C for 2 min and then increased to 300 °C
- 146 at $10 \,^{\circ}$ C min⁻¹. The metabolites were identified by retention time and mass spectra.
- 147

148 2.4. Simultaneous removal of dye and Cr(VI) by *C. elegans* biofilm and planktonic cells

- 149 *C. elegans* biofilm and planktonic cells were prepared as described in section 2.1 and
- incubated under the standard conditions with either MG (80μ M) or Cr(VI) (20 mg L^{-1}) only,
- 151 or a combination of both dye and metal. Decolourisation was monitored
- spectrophotometrically and Cr(VI) removal was assessed following the diphenyl carbazide
- 153 (DPC) method described by Maqbool et al. (2016). In order to test the reusability of the
- 154 biofilms for MG decolourisation and/or Cr(VI) removal, the aqueous solutions of dye and/or
- metal were continuously replaced with the fresh solutions after >95% of the pollutants were
- eliminated from the supernatant. The biofilms were rejuvenated after every three cycles ofdecolourisation by replacing the supernatants with 50 mL of fresh sabouraud dextrose broth
- and incubating for up to 16 h under shaking (150 rpm) at 28°C.
- To evaluate the ability of *C. elegans* biofilms to decolourise other dyes (reactive
 black-5, acid orange-7, direct red-81 & brilliant blue G) concurrently with Cr(VI) removal,
 triplicate biofilms were separately incubated with aqueous solutions containing 20 mg L⁻¹ of
 Cr(VI) and 50 mg L⁻¹ one of the selected dyes. Triplicate un-inoculated controls were also

- incubated for each treatment. Decolourization of the dyes in the supernatant was monitored
 spectrophotometrically at 597 nm (reactive black-5), 485 nm (acid orange-7), 540 nm (direct
 red-81) and 595 nm (brilliant blue G).
- 166

167 2.3.1. Impact of initial Cr(VI) concentration on removal efficiency

- 168 Triplicate biofilms were incubated with aqueous solutions of MG (80µM) and varying
- 169 concentrations (20 mg L^{-1} , 40 mg L^{-1} , 60 mg L^{-1} , 80 mg L^{-1} , 100 mg L^{-1} , 150 mg L^{-1}) of
- 170 Cr(VI). Triplicate un-inoculated flasks for each treatment were also incubated as controls.
- 171 The decrease of both pollutants in the supernatants was measured as described previously.
- 172 The re-usability of biofilms following rejuvenations at varying initial Cr(VI) concentrations
- 173 was also evaluated.
- 174

175 2.4.2. Impact of NaCl and metal ions on simultaneous removal of MG and Cr(VI) by *C*. 176 *elegans* biofilm

- 177 Biofilms were incubated with aqueous solutions containing MG (80µM) and Cr(VI) (20 mg
- 178 L^{-1}) plus varying concentrations (up to 100 g L^{-1}) of NaCl. The presence of 20 mg L^{-1} various

metal ions $(Ag^+, Cu^{2+}, Zn^{2+}, Mn^{2+}, Ni^{2+}, Ba^{2+}, Fe^{3+})$ on simultaneous removal of both

pollutants was similarly investigated. Triplicate un-inoculated flasks for each treatment werealso incubated as controls.

182

183 **2.5 Statistical analysis**

The results are presented as means \pm standard deviation. The means were compared using Least Significance Difference (LSD) test after the analysis of variance (ANOVA) at $p \le 0.01$ using the software R (3.4.1).

187

188 **3. Results**

189 **3.1.** MG decolourisation by *C. elegans*

- 190 MG was decolourised by *C. elegans* biofilm and planktonic cultures. Over the first 6 h
- incubation in the biofilm culture, the colour in the supernatant had decreased by
- approximately 60 % and, after 24 h incubation, almost a complete (> 95%) removal of colour
- 193 was observed in the supernatant (data not shown). Upon the second addition of dye to the
- 194 flasks the supernatant and biomass were monitored spectrophotometrically at different time
- points (Fig 1). The dye was removed from the supernatant within 15 min by absorption to the
- biomass (Fig 1A); the colour in the biomass dissipated more slowly (Fig 1 B and C). The

biomass of both the biofilm and planktonic cultures gave almost a similar pattern of colour

removal from the supernatant and absorbance over the incubation period. Thus, rapid initial

decolourisation of the supernatant through biosorption was followed by a slower

200 biodegradation of the MG dye.

In order to study the impact of decreasing pH on decolourisation of MG by *C. elegans* biofilm, the decolourisation experiments were carried out at pH from 7 to 3. The cultures incubated at pH values from 4 to 7 were found to decolourise more than 95% of the initially added MG in the supernatants within the first 24 hours. However, only 80% decolourisation of the supernatant was observed in the same period with cultures at pH 3 (Supplemental Information). Furthermore, the time for complete decolourisation (i.e. supernatant and

207 biomass) of MG by the cultures increased as pH was lowered.

208

209 **3.2.** Assessment of biodegradation of MG by *C. elegans*

The biomass from biofilm cultures incubated with 80μ M malachite green was extracted with ethyl acetate and the extractable metabolites were analyzed by GC-MS. The GC-MS analysis

revealed the presence of leucomalachite green, *N*-demethylated metabolites, 4-

213 (dimethylamino) benzophenone and aminobenzophenone (Table 1). The presence of these

214 metabolites suggests a stepwise demethylation followed by oxidative cleavage as previously

suggested by Cha et al. (2001). Interestingly, upon subsequent dye addition to the biofilm, no

216 metabolites were detectable by GC-MS after 24 h incubation, indicating complete

- 217 biodegradation.
- 218

219 **3.3.** Semi-continuous biofilm-catalyzed simultaneous removal of MG and Cr(VI)

3.3.1. Simultaneous removal of MG and Cr(VI) by *C. elegans* biofilm and planktonic cells

222 C. elegans biofilm and planktonic cultures were tested for their potential not only to remove

223 MG and Cr(VI) individually but also for simultaneous removal of MG and Cr(VI) in the

same solution. The data are summarized in Table 2 and show that MG removal was

comparable in planktonic and biofilm cultures whether in the absence or presence of Cr (VI),

- with approx. 80 % decolourisation within 16 h and complete degradation within 22-26 h.
- 227 Planktonic cultures were more effective at Cr(VI) removal than biofilm, with 83 % removed
- in 16 h compared to 71 %, and a shorter time required for complete removal (22 h compared
- with 24 h). However, whereas the efficiency of biofilm was not significantly impacted with

- the combination of dye and metal, the removal of Cr (VI) in planktonic cultures after 16 h
 decreased noticeably compared with the cultures incubated with the metal only.
- One of the main potential advantages of employing the biofilm is the ease of reusability, which was demonstrated in these experiments, showing that complete (> 95 %) removal of dye and metal was possible for at least 19 repeated additions (Table 2). Planktonic cultures are more difficult to recycle, as a centrifugation (or filtration) step is
- 255 Thanktome cultures are more unneut to recycle, as a centifugation (or mitation) step is
- necessary, and the suspended cells have previously shown to cease functioning after approx.three cycles (Amadio et al., 2013).
- 237 tł 238

239 **3.3.2.** Impact of initial Cr(VI) concentration on biofilm efficiency

Varying initial concentrations of Cr(VI) had an impact on the simultaneous removal of MG 240 and Cr(VI) by the C. elegans biofilm (Figure 2). After 16 h incubation, over 90% of the 241 initially added MG was decolourized in the solutions containing Cr (VI) concentrations up to 242 $60 \text{ mg } \text{L}^{-1}$; however, higher concentrations of the metal resulted in a decrease in 243 decolourisation ability. The total amount of Cr (VI) removed within the same period in these 244 experiments increased from 0.91 mg, when an initial concentration of 20 mg L⁻¹ was used, up 245 to 1.95 mg when the initial Cr (VI) concentration was 60 mg L⁻¹. At higher concentrations 246 247 the removal progressively declines. Notably, increasing the initial concentrations of Cr(VI) also resulted in an increase in the time required for complete (>95%) simultaneous removal 248 of Cr(VI) and MG, and a decrease in number of cycles of complete simultaneous removal of 249 both the pollutants (Table 3). 250

251

252 **3.3.3. Effect of NaCl and other metals on biofilm efficiency**

- 253 Simultaneous removal of MG and Cr(VI) by the C. elegans biofilm was not substantially
- affected by NaCl concentrations of 20 g L^{-1} (Fig 3 and Table 3); however, at higher
- concentrations the removal efficiency after 16 h, the time required for complete removal and
- the number of cycles of complete dye/metal removal were all affected. Fig 4 shows the effect
- of a selection of metal ions (2 mM) on the simultaneous removal of dye and Cr(VI). Most of
- the metals tested inhibited the removal of both pollutants to some degree, although complete
- (>95 %) removal was still achieved within 40 h in all experiments.
- 260

3.4. Simultaneous removal of Cr(VI) and other dyes by *C. elegans* biofilm

- 262 The ability of *C. elegans* biofilm for simultaneous removal of Cr(VI) and other dyes
- 263 (Reactive Black 5, Acid Orange 7, Direct Red 81 and Brilliant Blue G) was also examined.

264 This biofilm showed a good potential for parallel removal of Cr(VI) and different dyes from

- the culture supernatant (Table 4). After 40 h incubation, a complete (>95%) removal of
- 266 Cr(VI) was observed along with at least 85 % simultaneous removal of the initially added
- 267 dye. The dyes were biosorbed by the biofilms, but, unlike MG, the biomass was not
- completely decolourized, even after 120 hours incubation.
- 269

270 **4. Discussion**

271 Environmental pollution due to synthetic textile dyes is one of the leading contributors in 272 degradation of natural resource. This negative impact of synthetic dyes is intensified when these dyes loaded effluents are also accompanied by the presence of different pollutants. 273 Hexavalent chromium [Cr(VI)] is one of such pollutants which has often been found to co-274 exist as a contaminant with synthetic dyes in textile and tanneries effluents. Hence, there is 275 need to devise the strategies for concurrent removal of such co-existing pollutants and the 276 present study was conducted to evaluate the potential of *C. elegans* biofilm for simultaneous 277 removal of a synthetic dye, malachite green (MG), and Cr(VI) in a semi-continuous system. 278

279 The decolourisation of MG in planktonic and biofilm cultures occurred following a similar pattern, with the dye rapidly adsorbed by the biomass, followed by a slower 280 281 biodegradation step, resulting in complete removal of 80 µM dye in 24 h. This pattern of decolourisation with initial biosorption followed by degradation has also been observed in 282 other studies that focused on fungal biodegradation of MG (Jadhav and Govindwar, 2006; 283 Jasinska et al., 2012). It was observed here that the time taken for dye decolourisation by C. 284 285 elegans biofilm upon initial addition of MG was longer compared to the subsequent rounds, in contrast to planktonic cultures. One possible reason for this difference is that in the 286 biofilm there are specific genes required for decolourisation that are induced upon dye 287 addition, but in planktonic cells the genes are already expressed. Transcriptomic and 288 289 proteomic analyses of other fungi demonstrate that expression of genes can vary between planktonic and biofilm cultures (Gutierrez-Correa et al., 2012). 290

In general biofilms are stable and active over long periods (Halan et al., 2012) and *C. elegans* biofilms have been shown to be conveniently reused for biotransformations (Amadio et al., 2013; Quinn et al., 2015), thus they have potential for application in continuous or semi-continuous processes. The *C. elegans* biofilm can decolourize MG over a range of acidic pH, which is an added advantage; however, at pH 3 decolourisation ability is compromised. This is comparable with the decolourisation activity of some other fungal strain including *Penicillium ochrochloron*, which is completely inhibited at pH 3 (Shedbalkarand Jadhav, 2011).

Leucomalachite green, demethylated leucomalachite green, 4-(dimethylamino) 299 benzophenone and aminobenzophenone were observed as intermediate metabolites during 300 initial decolourisation of MG by C. elegans biofilm. Cha et al. (2001) identified mono-, di-, 301 302 and tri-demethylated derivatives of malachite green and leucomalachite green after decolourisation by suspended C. elegans ATCC 36112. The other metabolites detected in the 303 304 present study had not previously been identified from the fungus, but are known 305 intermediates in the biodegradation of malachite green in other microorganisms. For example, 4-(dimethylamino) benzophenone and 4-aminobenzophenone were observed during 306 degradation of MG by Micrococcus sp. strain BD15 (Du et al., 2013). 4-(Dimethylamino) 307 benzophenone has also been detected during decolourisation of MG by Shewanella 308 decolourationis NTOU1 under anaerobic conditions (Chen et al., 2010). 309

310 C. elegans biofilm as well as planktonic cultures were shown here for the first time to simultaneously remove MG and Cr(VI) from contaminated water. Although bacterial 311 312 cultures have been reported to concurrently remove Cr(VI) and different azo-dyes (Magbool et al., 2016; Anwar et al., 2014; Mahmood et al., 2013), to the best of our knowledge, there is 313 314 no report of the simultaneous removal of MG and Cr(VI) by any single microbial strain. Furthermore, the biofilm can tolerate the presence of Cr(VI) up to 60 mg L⁻¹, and up to 40 g 315 316 L^{-1} NaCl, but is sensitive to higher concentrations of both, resulting in longer times for complete removal of dye/metal and a reduction in the number of times the biofilm can be re-317 used. Simultaneous removal of dye/metal by the biofilm is possible even in the presence of 318 metal ions, such as silver and copper, albeit at a slower rate. 319

There are numerous reports of immobilized fungi applied to the decolourisation of 320 dye-contaminated water (Couto, 2009), but only a handful of these concern MG, and none 321 322 that also involve Cr(VI) removal. Barapatre et al. (2017) reported MG decolourisation in Aspergillus flavus and demonstrated that immobilization on a number of inert materials, such 323 as polyurethane foam and clay brick, resulted in improved decolourisation compared with 324 suspended culture. However, no experiments to investigate the recycling of the immobilized 325 326 fungus were done. In the present study, C. elegans was immobilized as a biofilm, which enabled repeated use (at least 19 cycles of dye/metal removal), which is attractive for 327 bioremediation applications. Furthermore, a screen of other dyes demonstrated that the 328 biofilm could biosorb these also, thus expanding its potential for remediation of dye-329 330 contaminated water.

331 Conclusion 332 Based on the findings of this study, it can be concluded C. elegans biofilm might serve as 333 potential bioresource to devise the strategies for simultaneous removal of Cr(VI) and MG 334 even in the presence of NaCl and metal ions that are characteristically present in real textile 335 336 and tanneries effluents. Confirmation of the re-usability of this biofilm is an important feature for its potential use in wastewater treatment processes, which require continuous operation. 337 338 339 Acknowledgements 340 This work was supported by the Irish Research Council (SH) and the Environmental Protection Agency (LQ). 341 342 **Conflict of interest** 343 344 None 345 346 **References** Amadio, J., Casey, E., Murphy, C. D., 2013. Filamentous fungal biofilm for production of 347 348 human drug metabolites. Applied Microbiology and Biotechnology 97, 5955-5963. Anwar, F., Hussain, S., Ramzan, S., Hafeez, F., Arshad, M., Imran, M., Maqbool, Z., Abbas, 349 350 N., 2014. Characterization of reactive red-120 decolorizing bacterial strain Acinetobacter junii FA10 capable of simultaneous removal of azo dyes and hexavalent chromium. Water 351 Air and Soil Pollution 225, doi: 10.1007/s11270-014-2017-7. 352 Barapatre, A., Aadil, K. R., Jha, H., 2017. Biodegradation of malachite green by the 353 ligninolytic fungus Aspergillus flavus. CLEAN – Soil, Air, Water 45, 1600045. 354 Bilandzic, N., Varenina, I., Kolanovic, B. S., Oraic, D., Zrncic, S., 2012. Malachite green 355 residues in farmed fish in Croatia. Food Control 26, 393-396. 356 Cha, C. J., Doerge, D. R., Cerniglia, C. E., 2001. Biotransformation of malachite green by the 357 fungus Cunninghamella elegans. Applied and Environmental Microbiology 67, 4358-4360. 358 Chen, C. H., Chang, C. F., Liu, S. M., 2010. Partial degradation mechanisms of malachite 359 green and methyl violet b by Shewanella decolorationis NTOU1 under anaerobic conditions. 360 Journal of Hazardous Materials 177, 281-289. 361 Cheung, K. H., Gu, J. D., 2007. Mechanism of hexavalent chromium detoxification by 362 microorganisms and bioremediation application potential: A review. International 363 Biodeterioration & Biodegradation 59, 8-15. 364

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467 **Figure legends**

- 468 Fig 1. The decolourisation malachite green in *C. elegans*. (A) Absorbance (617 nm) of
- supernatants (S/N) and biomasss (BM) after 15 min incubation with the fungus. Error bars
- 470 represent standard deviation n=2. (**B**) Absorbance spectra of methanolic extracts of
- 471 planktonic biomass. (C) Absorbance spectra of methanolic extracts of biofilm biomass. The
- 472 slightly lower absorbance in biofilm reflects the effectiveness of the extraction method using
- 473 methanol
- 474 Fig 2. The effect of initial Cr (VI) concentration on the simultaneous removal of MG and Cr
- 475 (VI) after 16 h incubation with the fungus. The means of MG removal compared using Least
- 476 Significance Difference (LSD) test after the analysis of variance (ANOVA) at $p \le 0.01$ (LSD)
- 477 value=17.43). The mean values labelled by the same letter(s) are not significantly different.
- 478 Fig 3. The effect of NaCl concentration on simultaneous removal of MG and Cr (VI). The
- 479 means of MG removal and Cr(VI) removal compared using Least Significance Difference
- 480 (LSD) test after the analysis of variance (ANOVA) at $p \le 0.01$ (LSD value for MG
- removal=13.04, LSD value for Cr(VI) removal=17.42). The mean values within either
- response (MG removal or Cr(VI) removal) labelled by the same letter(s) are not significantlydifferent.
- 484 Fig 4. The effect of metal ions on the removal of MG (A) and Cr (VI) (B) by C. elegans. The
- 485 means of MG removal and Cr(VI) removal at varying time intervals compared using Least
- 486 Significance Difference (LSD) test after the analysis of variance (ANOVA) at $p \le 0.01$ (LSD)
- 487 value for MG removal after 8 h= 4.57, LSD value for MG removal after 24 h= 5.73, LSD
- 488 value for Cr(VI) removal after 8 h= 5.01). The mean values within either response (MG
- 489 removal or Cr(VI) removal) at a specific time labelled by the same letter(s) are not
- 490 significantly different. The unlabelled mean values for MG removal over 40 h and Cr(VI)
- 491 removal over 24 hours were found statistically non-significantly different among themselves.
- 492
- 493

Intermediate products	Molecular structure	$T_R(min)$	m/z of M
Leucomalachite green	H ₃ C _N CH ₃ CH ₃ CCH ₃	14.65	330
Desmethyl Leucomalachite green	H ₃ C _N CH ₃ CH ₃	14.29	316
4-(Dimethylamino) benzophenone	O N(CH ₃) ₂	15.39	225
4-Aminobenzophenone	NH ₂	14.16	197

494 Table 1. GC-MS data for the metabolites of malachite green incubated with *Cunninghamella*495 *elegans* biofilm

- 498 Table 2. Removal of Cr(VI) and malachite green (MG) by the suspended cells (planktonic)
- and biofilm of *Cunninghamella elegans*. > 95 % Decrease in dye/metal is considered

Culture		Cr (VI)	MG	Simultaneous removal	
condition		Removal	Removal	Cr (VI)	MG
Planktonic	% Removal after 16 hours	82.9±5.9	81.6±4.4	74.3±3.4	79.5±6.1
	Time for complete removal (h)	22	22	24	24
Biofilm	% Removal after 16 hours	71.4±3.5	82.5±6.2	69.2±2.9	80.5±4.6
	Time for complete removal (h)	24	22	26	26
	No. of cycles of complete removal	24	19	23	20

500 complete removal.

501

502

Table 3. Effect of NaCl and initial Cr(VI) concentrations on simultaneous removal of Cr(VI) and malachite *cunninghamella elegans* biofilm. > 95 % Decrease in dye/metal is considered complete removal.

NaCl concentration	No NaCl		20 g L ⁻¹		40 g L ⁻¹		60 g L ⁻¹		80
	MG	Cr(VI)	MG	Cr(VI)	MG	Cr(VI)	MG	Cr(VI)	M
Time for complete removal (h)	24	24	24	24	28	32	32	54	96
No. of cycles of complete removal	>10	>10	>10	>10	6	4	3	2	1
Initial Cr(VI) concentration	20 mg L ⁻¹		40 mg L ⁻¹		60 mg L ⁻¹		80 mg L ⁻¹		10
	MG	Cr(VI)	MG	Cr(VI)	MG	Cr(VI)	MG	Cr(VI)	M
Time for complete removal (h)	24	24	24	32	28	52	48	72	76
No. of cycles of complete removal	>10	>10	6	5	2	2	1	1	1

	Colour Removal (%)		Cr(VI) Remo	val (%)
	24 h	40 h	24 h	40 h
Reactive Black 5	38.6 ± 3.6	90.7 ± 3.1	85.3 ± 4.5	98.6 ± 2.1
Acid Orange 7	51.9 ± 3.1	92.6 ± 3.9	82.1 ± 3.4	96.6 ± 2.6
Direct Red 81	72.7 ± 1.4	96.3 ± 2.6	89.1 ± 2.4	96.7 ± 3.1
Brilliant Blue G	58.9 ± 2.3	86.6 ± 3.9	90.1 ± 1.9	96.6 ± 2.8

Table 4. Simultaneous removal of Cr(VI) and various dyes by *Cunninghamella elegans* biofilm.

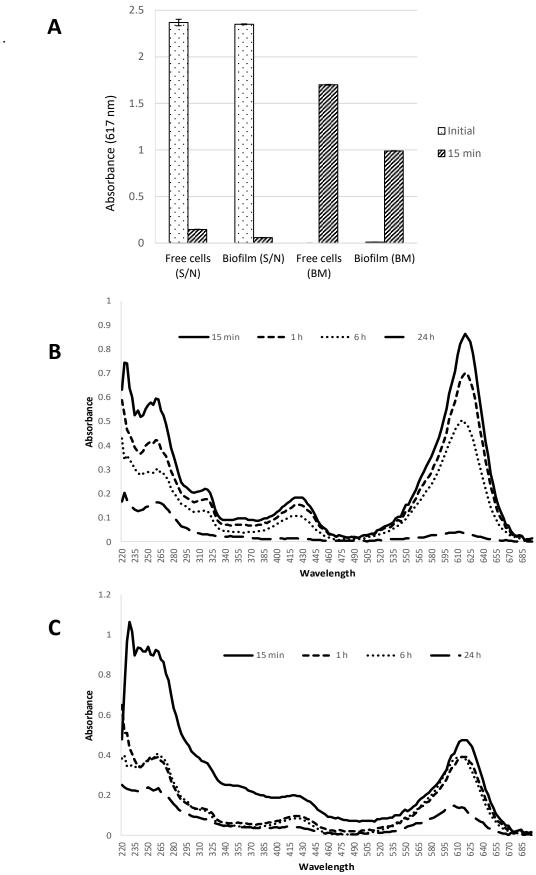
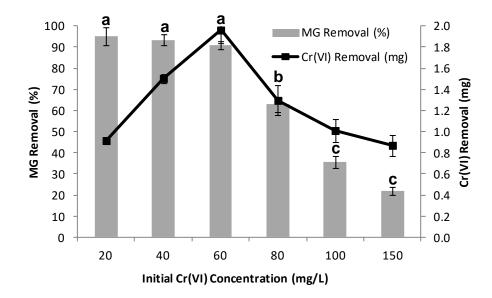


Fig 1.





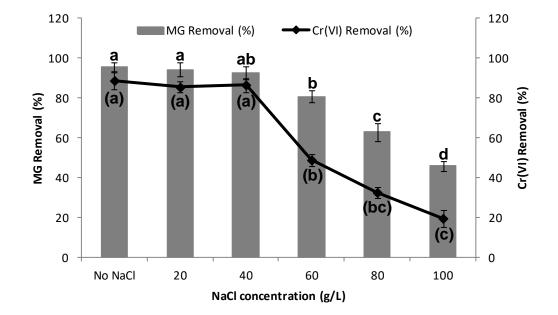
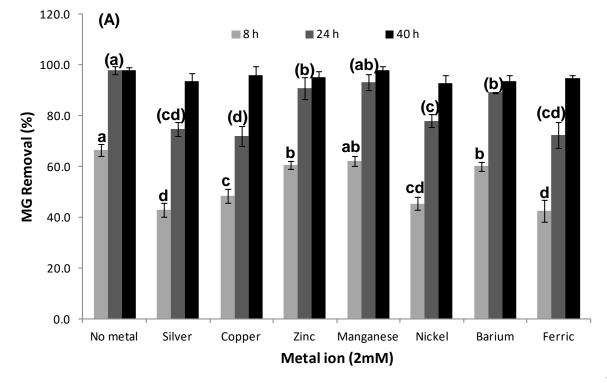


Fig 3





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